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Soil quality effects of compost and manure in arable cropping

Results from using
soil improvers
for 17 years in the
MAC trial

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I N S T I T U T E

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Soil quality effects of compost and manure in arable cropping - Results from using soil improvers for 17 years in the MAC trial

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Preface

Effects of organic applications on soil fertility and yield stability were evaluated. The evaluation provides knowledge of the different returns and yield stability strategies in relation to the soil quality parameters and ecosystem services.

In this research, long-term effects are evaluated of eight fertilization strategies based on compost, organic manure and mineral fertilizer and their impact on soil chemical, physical and biological parameters indicating overall soil fertility. The trial, started in 1999, is carried out at the arable farm Arenosa in Lelystad, The Netherlands. The 'Manure as a Chance' trial as it is called, is unique in its kind and, as far as known, there is no comparable experimental test in the world where the medium term effect of soil fertility can be assessed with thirteen different manure and compost types. This trial field has been set up and is managed by the Louis Bolk Institute.

We thank very much the farmer Jan van Geffen for his time and effort in managing the field. Financial support through the Dutch public-private program on Improving Soil Management, funded by the Dutch Ministry of Economic Affairs.

This evaluation could not be performed without the contributions of former colleague and founder of the trial Jan Bokhorst and the contribution of many colleagues and students in the course of the experiment.

We also thank Monique Hospers, Riekje Bruinenberg en Hans Dullaert of the Louis Bolk Institute for their contribution to the field and lab work, the laboratory of Wageningen Environmental Research for analyzing soil biological parameters and Harm Keidel (LIOS) for determining nematode compositions. Bart Timmermans contributed with inspiring discussions on the results.

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Summary

There are many factors to consider when deciding how to maintain a sustainable soil fertility in arable cropping. Multi-year research into the divergent effects of fertilizers can help considerably. The hypothesis of this study is that the fertilizer strategy including the quantity and quality of the organic substance applied, in the end determines soil quality as well as the yield capacity of the soil.

In this study three fertilization strategies are compared: (i) fertilization mainly for the purpose of feeding plants: mineral fertilizer or slurry treatments; (ii) Fertilization with the purpose of feeding the soil: plant composts, household compost and nature compost treatments; (iii) Fertilization both for plant and soil feeding: poultry manure, household compost&slurry and deep stable manure treatments. North of the city of Lelystad, The Netherlands these eight fertilizer treatments are compared in a completely randomized block design in an intensive rotation including crops like potatoes, carrots, parsnip and salsify. Soil samples were taken after the growing season in 2016, 17 years after the start of the experiment.

Yields for the plant feeding, soil & plant feeding and soil feeding strategies diverged during this period. The relative yield of the plant feeding strategy remained stable. For the soil feeding strategies, yields declined over time, while the soil & plant feeding fertilization strategies resulted in increasing yields during the course of the experiment.

After 17 years, significant effects were found for organic matter and labile organic matter levels in the soil. Highest levels of organic matter (+41%) were found in the nature compost treatment. Also deep stable manure (+31%) and household compost&slurry (+22%) increased soil organic matter levels significantly. Mineral fertilizer and slurry remained at levels similar to the start of the experiment (1.6%). Higher soil organic matter additions corresponded with higher soil organic matter levels found after 17 years. Labile soil organic matter (determined as Hot Water extractable Carbon) was found to be highest in the stable manure, followed by household compost&slurry and nature compost treatments.

Mineralization of nitrogen in the soil was affected by the fertilizer additions. It remained lowest in the mineral fertilizer treatment. The deep stable manure resulted in a mineralization, which was considerably higher if compared to mineral fertilizer (+ 70%) followed by nature compost (+ 58%), and the household compost&slurry combination (+39%). The soil mineralizable C followed a pattern similar to the level of soil organic matter.

Limited effects of fertilizer additions were found on soil organisms. Earthworms showed highest numbers in the poultry manure treatment compared to all other treatments. Organic matter additions did not change overall bacterial biomass, but fungal biomass was significantly higher in the mineral fertilizer treatment. The low mineralization in the mineral fertilizer treatment combined with a high fungal biomass could be an indication of a low nutrient availability at the end of the growing season.

Nematode abundance was highest in the deep stable manure and nature compost treatments, followed by the mineral fertilizer and plant compost treatments. Herbivorous and bacterivorous nematodes dominated, with highest numbers in the nature compost followed by mineral fertilizers and plant compost treatments.

Overall, it might be concluded that best results are obtained with fertilizers used for both crop nutrition and building soil quality: deep stable manure, household compost&slurry and poultry manure. Although there are positive indications on total organic matter, total C and labile organic matter, the build-up of the soil organisms is slow which might also be a result from the light soil and the intensive rotation and soil cultivation.

Samenvatting

Bij het inzetten op een duurzame bodemvruchtbaarheid voor de akkerbouw moet met meerdere factoren rekening worden gehouden. Meerjarig onderzoek naar de uiteenlopende effecten van meststoffen kan aanzienlijk helpen om inzicht te krijgen in de effecten van meststoffen voor gewas, de bodem en het bodemleven. In deze studie is de hypothese dat de hoeveelheid en kwaliteit van de toegepaste organische meststoffen op de lange termijn bepalend is voor de bodemkwaliteit en de opbrengstcapaciteit van de bodem.

Drie bemestingsstrategieën zijn daarbij vergeleken: (i) bemesting voornamelijk gericht op het voeden van de planten: minerale meststof en drijfmest; (ii) Bemesting voornamelijk gericht op het voeden van de bodem: groencompost, GFT-compost en natuurcompost; (iii) Bemesting gericht op het voeden van zowel de planten- alsook de bodem: pluimveemest, een combinatie van GFT&drijfmest en potstalmest. In een gewarde blokkenproef op het biologische bedrijf Arenosa ten noorden van Lelystad zijn de effecten van bovenstaande acht bemestingsvarianten in een intensieve rotatie met o.a. aardappel, peen, pastinaak en schorseneer na 17 jaar met elkaar vergeleken. De opbrengsten van gewassen lopen voor de 3 strategieën uiteen. De relatieve opbrengst van de op de voeding van planten gerichte strategie bleef stabiel. Bij de bodem voedende strategie daalden de opbrengsten over de jaren, terwijl de bodem- en planten voedende bemestingsstrategie resulteerde in toenemende opbrengsten in de loop van het experiment.

De grootste toename aan bodemorganische stof werd in de natuurcompost gevonden (+ 41%). Ook de potstalmest resulteerde in een toename (+31%) evenals de GFT&drijfmest variant (+ 22%). Bij de minerale mest en drijfmest bleef het organische stofgehalte dicht bij de uitgangssituatie (1,6%). De in de bodem aangetroffen organische stof correleerde sterk met de totale hoeveelheid organische stof toegediend met de bemesting gedurende 17 jaar. De meeste labiele organische stof werd aangetroffen in de stalmest variant, gevolgd door GFT compost&drijfmest en natuurcompost.

De mineraliseerbare stikstof in de bodem werd sterk beïnvloed door de toegediende meststoffen. Mineraliseerbare stikstof bleef het laagst in de minerale mest variant. De potstalmest resulteerde in de hoogste mineralisatie (+ 70% ten opzichte van minerale mest) gevolgd door natuurcompost (+ 58%) en GFT compost&drijfmest (+ 39%). De mineraliseerbare koolstof C volgde een patroon dat vergelijkbaar is met het verloop van de organische stof in de bodem.

De effecten van de organische meststoffen op bodemleven bleven ook na 17 jaar relatief beperkt. Regenwormen vertoonden hogere aantallen in de pluimveemest vergeleken met alle andere varianten. De organische meststoffen hadden geen effect op de totale bacteriële biomassa, wel werden verschillen in schimmelbiomassa gevonden met significant meer schimmels in de minerale mest variant. De lage mineralisatie in deze variant in combinatie met de hoge schimmelbiomassa kan een aanwijzing zijn voor een lage beschikbaarheid van voedingsstoffen ten tijde van de meting in het najaar. Mogelijk is hierbij sprake van grotere hoeveelheden mycorrhiza schimmels.

Grootste aantallen nematoden werden in de potstalmest en natuurcompost gevonden gevolgd door de minerale mest en groencompost varianten. Herbivore en bacterivore nematoden domineerden bij de natuurcompost gevolgd door minerale mest en groencompost.

Al met al kan worden geconcludeerd dat de beste resultaten kunnen worden behaald met meststoffen die zowel op voeding van het gewas alsook op de opbouw van de bodem zijn gericht: potstalmest, GFT compost&drijfmest en pluimveemest. Zowel de organische stof alsook de labiele organisch stof laten daarbij duidelijke effecten zien, maar veranderingen in bodemleven komen minder naar voren en lijken traag te verlopen. De lichte zavelgrond maar ook het relatief intensieve bouwplan met een intensieve grondbewerking tot gevolg zouden hieraan ten grondslag kunnen liggen.

1 Introduction

The Dutch Agri & Food Agenda aims to make agricultural production systems more circular. For instance by reducing and recycling waste: 'The growth of the world's population demands additional efforts to improve productivity and resource efficiency. Robust plant production with a high soil quality is central for achieving this aim.

However, for farmers legislation is becoming more complex which results in fewer options for improving yields and at the same time maintaining soil fertility. There are different requirements to the use of various natural inputs like manure or compost (levels of nitrogen, phosphate, etc.). Reduction of nutrient losses like nitrate leaching is essential within the European nitrate directive. This often makes it easier to meet requirements with mineral fertilizers because it is available in any desired composition. Organic fertilizers however, contribute to a robust production through the unique combination of nutrients with carbon rich materials. This means nutrition for soil life with biodiversity and nutrient availability as a result. In addition, carbon storage and increased water regulation of soils are societal issues that require more and more focus and attention of farmers when using inputs.

How to make the right choices between short-term yield and long-term production stability, between business interests and societal requirements? Are the apparent contradictions bridgeable here? In order to combine and optimize all requirements, knowledge about the fate of organic matter is required. The effect of organic matter inputs on soil quality is only partly visible in the medium to long term.

There are many factors to consider when deciding how to fertilize agricultural crops. Each fertilizer has its advantages and disadvantages, and they are not easy to weigh up. Multi-year research into the divergent effects of fertilizers can help considerably. In the first instance, it is important to know how to achieve good yields. The next question is whether the choice of fertilizer based on this premise will continue to give a good yield over the longer term. Think, for example, of the build-up of organic matter in the soil. There may be a link between fertilizer use, soil biodiversity and disease suppression; and this in turn affects production losses due to pests and diseases, and the use of crop protection products. In respect of climate change it is important to maintain, or (based on the field trial) even increase the organic matter content of the soil. Finally, biodiversity is increasingly considered as an important product in itself. With so many issues to consider, it is not easy to make the right choice. In this publication, we discuss the work done in the field trial and the most significant results after seventeen years of following different fertilization strategies.

The hypothesis of this study is that the fertilizer strategy including the quantity and quality of the organic substance applied, in the long run determines soil quality as well as the yield capacity of the soil. More specific questions are:

- What is the impact of the amount and composition of the manure or compost on soil quality in the long run?
- Which type and composition of organic matter enhance the required services like stability of production, soil biodiversity and carbon sequestration in the soil?
- Which fertilizer or combination of fertilizers performs best within legal requirements to serve these multiple goals?

2 Materials and methods

Study site and experimental design

The experiment is located north of the city of Lelystad in the province of Flevoland (52.32° N, 5.30° E). Since 1999, thirteen fertilization treatments are compared in a completely randomized block design with four replicates. Every field plot is 7x9 m.

The trial field has a calcium-rich loam soil with 9% clay and 4.4% lime. At the start of the experiment, the soil organic matter content was 1.6% (Table 1). The top soil is around 30 cm deep and deeper a stratified subsoil is found with alternating layers with low humus content.

Table 1: Soil parameters at the start of the experiment in 1999.

Soil depth cm	Organic matter %	pH-KCl	P-water mg P ₂ O ₅ l ⁻¹	P-Al mg P ₂ O ₅ 100 g ⁻¹	K-HCl mg K ₂ O 100 g ⁻¹
0-20	1.6	7.6	55	37	23

Thirteen types of fertilization are compared which can be divided in three types of fertilization strategy:

- Fertilizers used mainly for the purpose of crop nutrition: mineral fertilizer and slurry.
- Fertilizers used mainly to build up soil quality: plant composts, household compost and nature compost.
- Fertilizers used both for crop nutrition and building soil quality: poultry manure, household compost&slurry and deep stable manure.

During the experiment, manure and compost addition was limited by a maximum nitrogen mineralization of 100 kg N ha⁻¹ from fertilizer applications. As fertilizer were applied two years out of three, the average application of nitrogen mineralized from fertilization is 67 kg N ha⁻¹ yr⁻¹. For some treatments, a nitrogen application of 100 kg leads to an exceedance of maximum allowed phosphorus fertilization. For these treatments, a maximum of 120 kg P₂O₅ ha⁻¹ was taken instead. With fertilizers applied two years in three, an average of 80 kg P₂O₅ ha⁻¹ year⁻¹ was applied which is the maximum within the statutory norm at the start of the experiment. 120 kg P₂O₅ could also be applied in the case of nature compost (with low levels of heavy metals).

In the case of household compost and plant compost only the legally permissible level of 6000 kg dry matter per ha per year was applied which meant that the 100 kg nitrogen mineralized per ha and the 120 kg P₂O₅ per ha levels were not achieved in these treatments. As this did not constitute a complete fertilization programme, an extra treatment was added. This was a combination of household compost and cattle slurry, with a total of 100 kg N-mineralized per ha.

In case of the household and plant compost, considerably less nitrogen was applied than in the other treatments. In the treatments with little or no organic matter, such as mineral fertilizer and slurry, far less soil mineralization (i.e. from previous years' fertilizer applications) was built up than in, for example, the deep stable manure treatment. This must be taken into account when evaluating the results

For this evaluation, more frequent and elaborate measurements were done in eight of the thirteen treatments. In this study, the main focus is on these eight treatments; mineral fertilizer (M), slurry (S), poultry manure (PM), household compost&slurry (HCS), deep stable manure (DSM), plant compost (PM; lightly composted), household compost (HC) and nature compost (NC).



Illustration 1: Fertilizer application in the field (from top left to bottom right) mineral fertilizer (M), slurry (S), poultry manure (PM), household compost&slurry (HCS), deep stable manure (DSM), plant compost (PC), household compost (HC) and nature compost (NC).

Fertilization

Manure and compost addition was limited by a maximum nitrogen mineralization of 100 kg nitrogen per ha per year from the fertilizer applications. As fertilizer was applied two years in three, the average nitrogen mineralized from fertilization was 67 kg N ha⁻¹ mineralized or 80 kg P₂O₅ ha⁻¹. In order to do this, the amount of manure applied was based on measured N and P₂O₅ content of the manure. If no measurements were available beforehand, the amount was based on the nutrient contents of previous years. Because the nutrient content is not constant, applied nutrient application varied somewhat over years. The manures were applied in May, regardless of the preceding crops.

Table 2 summarizes the average nutrient and organic matter additions of the eight selected treatments. Table 3 summarizes the years of application.

Table 2: Fertilizer treatments compared in this study with the strategy and realised applications of nitrogen, phosphate and organic matter during the period 1999-2016.

Treatment	Strategy	Set-up application	Total N kg N ha ⁻¹ year ⁻¹	Phosphate kg P ₂ O ₅ ha ⁻¹ year ⁻¹	Dry matter kg d.m. ha ⁻¹ year ⁻¹	Organic matter kg o.m. ha ⁻¹ year ⁻¹
Mineral fertilizer (M)	Plant	67 kg N mineralized	66	55	551	0
Slurry (S)	Plant	67 kg N mineralized	99	38	2027	1426
Poultry manure (PM)	Soil & plant	80 kg P ₂ O ₅	89	79	2169	1601
Household compost & slurry (HCS)	Soil & plant	67 kg N mineralized	156	74	7102	2852
Deep stable manure (DSM)	Soil & plant	67 kg N mineralized	167	85	6746	4533
Plant compost (PC)	Soil	6000 kg d.m.	49	26	4911	1426
Household compost (HC)	Soil	6000 kg d.m.	65	39	5215	1525
Nature compost (NC)	Soil	80 kg P ₂ O ₅	175	79	19281	6149

Crop rotation

Activities at the field site were integrated in the regular farming practices and crop rotation of the organic farm. Several crops were grown in the trial over years. Table 3 summarizes the crop rotation over time. In the winters of 2011, 2012, 2013 and 2015, rye was sown as a cover crop.

Table 3: Field crops grown in the trial during the experimental period. Years of fertilizer additions are indicated.

Year	Crop	Fertilization	Cover crop
1999	Red cabbage	Yes	
2000	Potatoes	Yes	
2001	Beetroot	Yes	
2002	Carrots	No	
2003	Parsnip	Yes	
2004	Broccoli	No	
2005	Pumpkin	Yes	
2006	Cauliflower	Yes	
2007	Potatoes	No	
2008	Salsify	Yes	
2009	Parsnip	Yes	
2010	Pumpkin	No	
2011	Parsnip	Yes	Rye
2012	Potatoes	Yes	Rye
2013	Salsify	No	Rye
2014	Leek	Yes	
2015	Sweetcorn	Yes	Rye
2016	Salsify	No	

Crop yields

Crop yields were determined annually by harvesting and weighing the crops. For row crops, two rows in the centre of each plot were harvested over a length of 1.5 m. The fresh weight was determined and samples were dried at 70 °C to determine dry weight.

In order to make yields comparable over years with different crops, the relative yield was calculated for each treatment. The relative yield for a given year was calculated as the yield relative to the average yield over all treatments in that year.

$$Rel. yield_{it} = \frac{Y_{it}}{\frac{1}{n} \sum_{i=1}^n Y_{it}} * 100$$

Where t is the year the average is calculated for and Y_i is the yield for treatment i .

Soil sampling and chemical, physical and biological parameters determined

Soil samples were collected for chemical, physical and biological parameters on 15th September 2016. A bulk sample of 40 cores (0-10 cm, diameter of 2.3 cm) per plot was collected, sieved through a 1 cm mesh, homogenized and stored at field moisture content at 4 °C before analysis. Subsamples were taken for chemical, microbiological and nematode analysis.

Table 4: Soil chemical, physical and biological parameters determined in this study

Parameter	Depth (cm)	Unit
pH-KCL	0-10	-log[H ⁺]
Soil Organic Matter	0-10	%
Hot Water extractable C	0-10	µC g ⁻¹
C-total	0-10	g C kg dry soil ⁻¹
N-total	0-10	g N kg dry soil ⁻¹
P-total	0-10	mg P ₂ O ₅ 100 g ⁻¹
P-plant	0-10	mg P kg ⁻¹
K	0-10	mg K kg ⁻¹
Ca	0-10	mmol Ca L ⁻¹
Mg	0-10	mg Mg kg ⁻¹
Bulk density	2,5-7,5	g cm ⁻³
Soil structure	0-10	% crumb, sub angular, angular
Soil structure	10-20	% crumb, sub angular, angular
Macro pores	0-10	n m ⁻²
Macro pores	10-20	n m ⁻²
Mineralizable N	0-10	mg N kg dry soil ⁻¹
Mineralizable C	0-10	mg CO ₂ kg dry soil ⁻¹
Number of earthworms	10-20	n m ⁻²
Bacterial biomass	0-10	µg C g dry soil ⁻¹
Fungal biomass	0-10	µg C g dry soil ⁻¹
Active fungi	0-10	µg C g dry soil ⁻¹
Number of nematodes	0-10	n 100 g soil ⁻¹

Chemical parameters

For each site, soil from the bulk sample was oven-dried at 40 °C prior to analysis of soil acidity (pH-KCl), soil organic matter (SOM), total carbon (C-total), total nitrogen (N-total), total phosphorous (P-total) and CaCl₂-extractable P (P-plant). All chemical analysis were determined at Eurofins Agro. Soil pH-KCl was measured in 1 M KCl. Soil organic matter was calculated from determination of organic C by near infrared spectroscopy (NIRS).

C-total was measured by incineration of dry material at 1150 °C, after which the CO₂ produced was determined by an infrared detector (LECO Corporation, St. Joseph, Mich., USA). Amount of non-organic C was negligible. For determination of N-total, evolved gasses after incineration were re-

duced to N₂ and measured with a thermal-conductivity detector (LECO Corporation, St. Joseph, Mich., USA).

Hot Water extractable Carbon (HWC) was determined as the amount of dissolved organic carbon that is released during incubation of a soil sample in hot water during 16 hours at 80°C (Ghani et al, 2003). This is a measure of easily decomposable (labile) organic carbon. The HWC fraction of organic matter is rich in amorphous polysaccharides (mucigel) which originate mainly from microbial exudates and to a lesser extent from plant exudates. This fraction is highly available to microorganisms and is also regarded as one of the key labile components of organic matter responsible for soil micro-aggregation which is an important soil physical parameter to consider in terms of soil quality (Ghani et al., 2003; Haynes, 2005).

P-total was measured with Fleishmann acid (Houba et al., 1997) and plant available P, K, Ca and Mg in a 0.01 M CaCl₂ extract according to Houba et al. (1997).

Physical parameters

To determine soil bulk density, stainless steel rings were inserted at 2.5-7.5 cm soil depth. In each of the four replicate plots, four density samples were taken of 100 cm³. After drying (24 h at 105°C) the weight per volume was determined.

Soil structure was determined at each plot in 1 block (20 cm x 20 cm x 10 cm) in the soil layers 0-10 cm and 10-20 cm. Soil of this block was divided by visual observation into crumbs, sub-angular blocky elements (sub-angular) and angular blocky elements (angular). These were expressed as a percentage of total soil volume in the block according to Koopmans et al. (2009).

On the horizontal surfaces (20 cm x 20 cm) exposed at 10 cm and 20 cm depth, the total number of macro pores or earthworm burrows with a diameter of > 2 mm were counted.

Biological parameters

Mineralizable N was determined by anaerobic incubation of 16 g of soil in 40 ml water for 1 week at 40 °C (Keeney and Nelson 1982; Canali and Benedetti 2006). These incubation conditions are optimal for quick mineralization of organic matter by anaerobic bacteria. The lack of oxygen prevents conversion of released NH₄⁺ to NO₃⁻ (nitrification) and uncontrolled N losses by denitrification cannot occur. After 1 week, NH₄⁺ contents were determined by Segmented Flow Analysis.

Mineralizable N is a measure of labile, potentially plant available, organic N.

Mineralizable carbon was measured as CO₂ production in three replicates per plot. Forty grams of rewetted soil at 70% WHC was placed in a 0.5 L jar along with 10ml of 1M KOH and incubated for one week at 20°C (Anderson, 1982). After one week, unreacted alkali in the KOH traps was back-titrated with a 1 N HCl solution to determine CO₂-C. Mineralizable C was calculated for the 7-days period and is a measure of labile C that can be decomposed in the short term.

Earthworms were sampled in two blocks (20 cm x 20 cm x 20 cm) per plot. The earthworms were hand-sorted in the field, counted and fixed in alcohol prior to identification. Numbers are expressed per m². Adults were identified according to species. A distinction was made between epigeic species (pigmented, living superficially in the litter layer, little burrowing activity, (2) endogeic species (living in burrows at approximately 10-15 cm depth) and (3) anecic species (relatively large worms, living in vertical burrows) (Bouché, 1977). Since only endogeic species were found in the trial field all results refer to endogeic species.

Bacteria and fungi were determined in soil samples of 0-10 cm depth. Bacteria were measured by confocal laser scanning microscopy and automatic image analysis (Leica TCS SP2 and Qwin Pro), after staining of soil smears with DTAF. This is a fluorescent dye which binds to proteins (Bloem et al., 1995). From the number and cell volume, bacterial biomass was calculated and expressed as $\mu\text{g C g}^{-1}$ soil. Fungi in soil smears were stained with Differential Fluorescent Stain, a mixture of two stains: fluorescent brightener (blue) which binds to cell walls (polysaccharides) and europium chelate (red) which binds to nucleic acids (DNA and RNA) (Morris et al., 1997). Thus active and inactive hyphae were distinguished. In addition unstained melanized hyphae were counted by switching to transmitted light. The total hyphal length measured under the microscope was used to calculate fungal biomass in terms of $\mu\text{g C g}^{-1}$ soil (Bloem and Vos, 2004).

Nematodes were determined from a 100 ml sub-sample from which the free-living nematodes were extracted using the Oostenbrink elutriator (Oostenbrink, 1960). Total numbers were counted and expressed per 100 g fresh soil. Nematodes were fixed in hot formaldehyde 4% and at least 150 randomly selected nematodes from each sample were identified to genus and species whenever possible. Nematode genus and species were assigned to trophic groups following Yeates et al. (1993) and allocated to the colonizer-persister groups (cp-groups) following Bongers (1990) and Bongers et al. (1995). The Maturity Index was calculated as the weighted mean of the individual cp-values, in accordance with Bongers (1990). It is an ecological measure, which indicates the condition of an ecosystem based on nematode species composition.

Statistical analysis

The data were analysed with GENSTAT (13.3th edition, VSN International Ltd., Hemel Hempstead, UK) using the General Linear Models procedure in Genstat. Treatment and replicates were used as factors in the model. Significant differences were determined based on least significant differences (LSD, $p < 0.05$).

3 Results

Crop yield and development

Crop yields were significantly influenced by fertilizer application (Figure 1). Comparing the average yield of salsify in 2016 shows that the fertilization with mineral fertilizer or plant compost resulted in lowest yields compared to the other fertilization treatments. Highest yields were found in the poultry manure, household compost&slurry and deep stable manure treatments (Figure 1) all belonging to the soil & plant feeding strategy.

Over a timeframe of 17 years, yields for the plant feeding, soil & plant feeding and soil feeding strategies diverged. The relative yield of the plant feeding strategy was almost stable over the period of 17 years, but yields varied between different years, crops and treatments. For the soil feeding strategies, yields declined over time, while the soil & plant feeding fertilization strategies resulted in increasing yields during the course of the experiment (Figure 2). All changes are relative to the average yield of a given year, therefore absolute changes cannot be determined from these results.

Variability in yield between years was large, for example nature compost showed a yield 2% above average in 2014 (leek) and 56% below average in 2015 (sweetcorn). Of the compost strategies feeding the soil, plant compost showed the lowest yields overall years and household compost the highest, though this difference was not significant. For the plant & soil feeding fertilizers the highest yield was found for the household compost&slurry treatment.

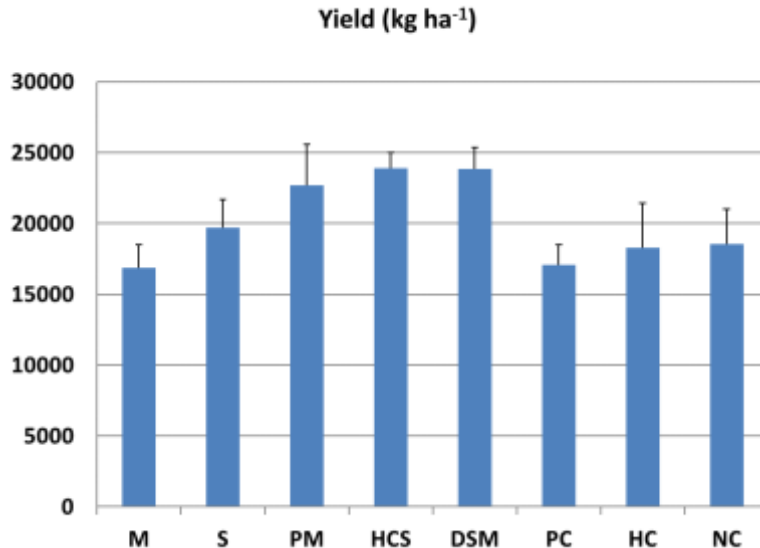


Figure 1: Effect of fertilization treatment on fresh crop yield of salsify in 2016 for mineral fertilizer (M), slurry (S), poultry manure (PM), household compost&slurry (HCS), deep stable manure (DSM), plant compost (PC), household compost (HC) and nature compost (NC).

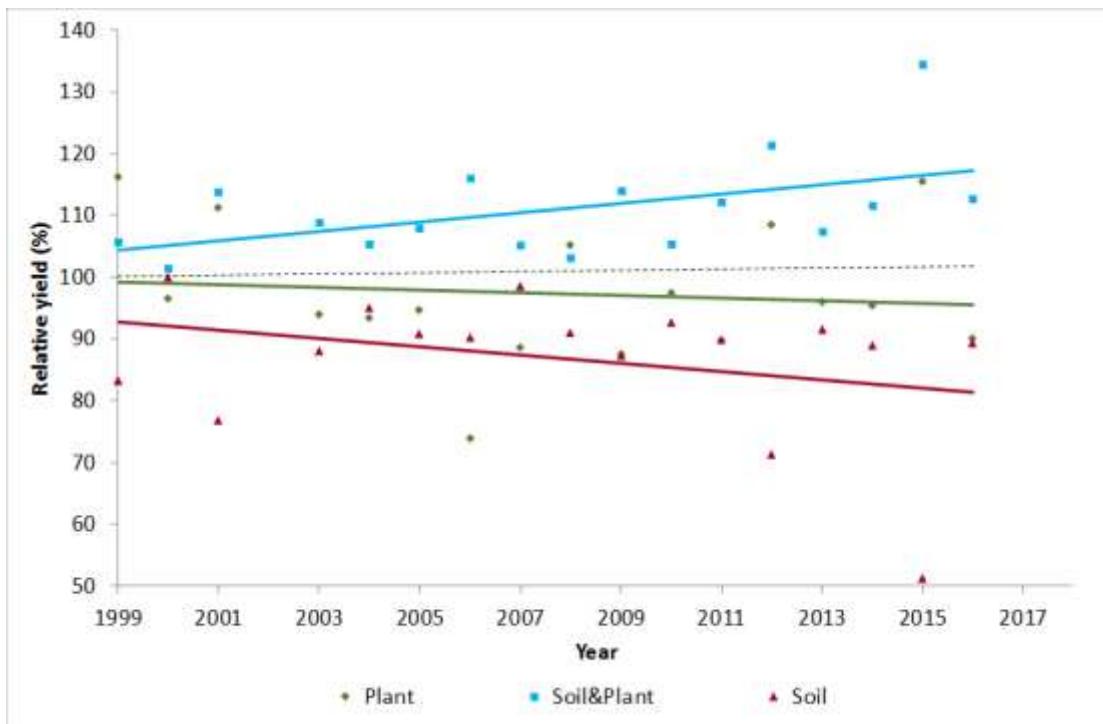


Figure 2: Changes in average relative yield over time for all strategies: plant feeding, soil & plant feeding and soil feeding.

Soil chemical parameters

Figure 3 indicates soil organic matter levels differed significantly between treatments ($p < 0.001$), seventeen years after treatments started. Highest levels of organic matter (2.3%; +41%) were found in the nature compost treatment. Also deep stable manure (+31) and household compost&slurry (+22%) increased organic matter levels significantly. Mineral fertilizer and slurry remained with 1.7% soil organic matter at levels similar to the start of the experiment (1.6%). Other treatments with poultry manure, plant compost and household compost showed a slight tendency to increase but differences with the start, mineral and slurry were not significant (Appendix 1). Organic matter levels seemed to be the result of the amount of organic matter applied (Table 2), since higher additions correspond with higher soil organic matter levels found after 17 years. It is striking that organic matter levels in the mineral treatment did not descend but remained at levels similar to the start of the experiment.

The hot water extractable carbon varied between 277 and 377 $\mu\text{C g}^{-1}$ dry soil. It was highest with deep stable manure, followed by household compost&slurry and nature compost. The hot water extractable carbon was found to be relatively low in the mineral fertilizer, slurry and poultry manure treatments ($p = 0.003$). But also in plant compost the hot water extractable carbon remained fairly low (Figure 3). Although with nature compost the highest amount of organic matter was given, the similar hot water extractable carbon as compared to deep stable manure and household compost & slurry indicates a lower decomposition rate of the nature compost.

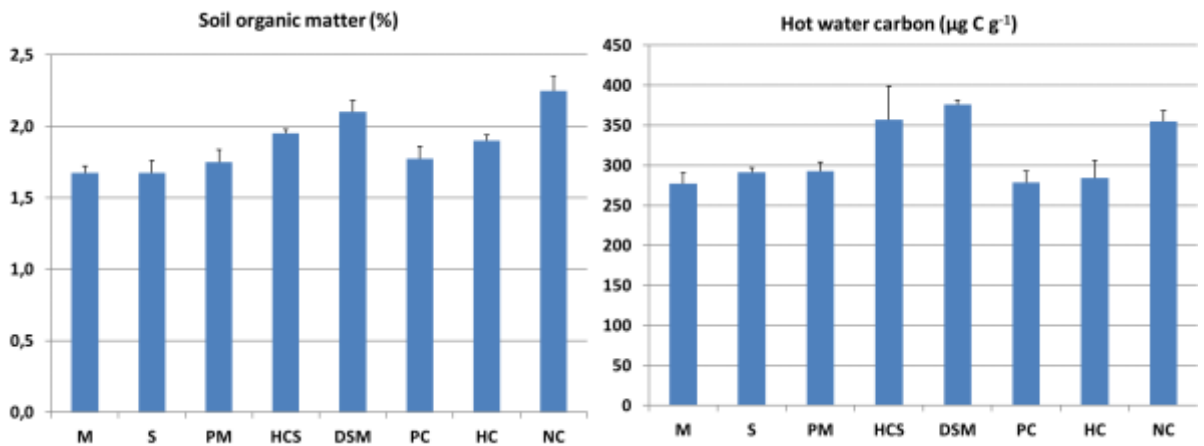


Figure 3: Effect of fertilization treatment on organic matter in the soil (left) and hot water carbon (HWC, right) for mineral fertilizer (M), slurry (S), poultry manure (PM), household compost&slurry (HCS), deep stable manure (DSM), plant compost (PC), household compost (HC) and nature compost (NC) treatments.

Differences between treatments became more pronounced since 2006 but average levels remained similar to 2006. In 2006, HWC with nature compost was higher than with plant compost and mineral fertilizer, but no other significant differences were observed (Koopmans et al., 2006).

Other chemical soil parameters are presented in Appendix 1. A slightly significant effect of the fertilizer additions on pH-KCL was found after 17 years of fertilizer additions with somewhat higher levels in the plant and household compost treatments compared to the deep stable manure and nature compost treatments. C-total and N-total also significantly differed between the fertilization treatments. Highest levels for C- and N-total were found in the nature compost treatments followed by deep stable manure and household compost&slurry. No differences were found in P-total. Highest levels of plant available P were found in deep stable manure followed by household compost&slurry and poultry manure. This corresponds with treatments receiving annual levels of P₂O₅ of about 80 kg ha⁻¹. An exception is the nature compost; here high P₂O₅ additions did not result in higher availability. In addition, K and Mg showed highest availability in the deep stable manure treatment followed by the household compost&slurry. No significant effect of treatments on Ca availability was found.

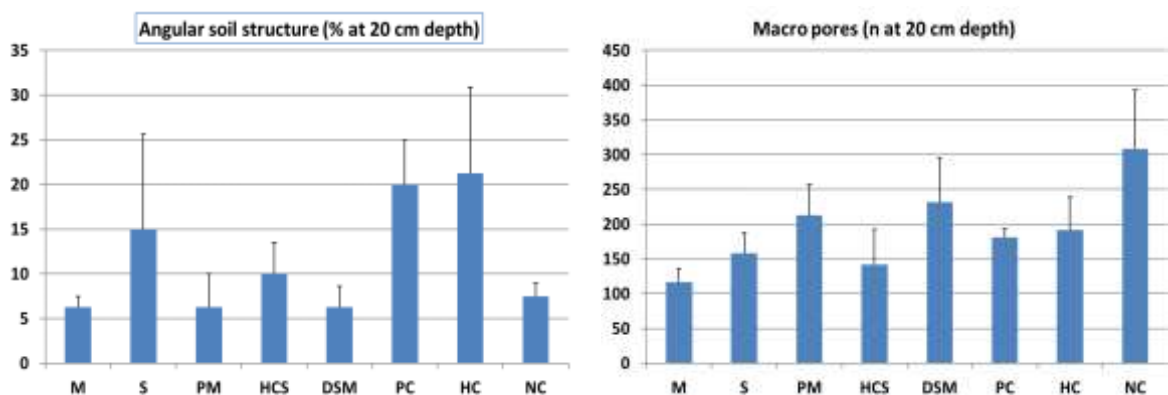


Figure 4: Effect of fertilization treatment on angular soil structure (left) and macro pores at 20 cm soil depth (right) for mineral fertilizer (M), slurry (S), poultry manure (PM), household compost&slurry (HCS), deep stable manure (DSM), plant compost (PC), household compost (HC) and nature compost (NC) treatments.

Soil physical parameters

Fertilizer treatments did not significantly influence soil physical parameters in the upper 0-20 cm soil layer. This may be partly due to the intensive soil cultivation at the site, which erases physical changes easily. Bulk density was 1.25 g cm^{-3} . Soil structure at 0-10 and 10-20 cm depth indicated no differences between treatments and a very high within plot variability in angular soil structure. Macro pores, mainly earthworm burrows, were found at numbers of 405 and 193 m^{-2} at 10 and 20 cm soil depth respectively, indicating relatively high levels for an intensive vegetable rotation. However, no statistical significant differences could be found as a result of, again, high variability.

Soil biological parameters

In the autumn of 2016, the nitrogen mineralization was on average 15.7 mg kg^{-1} dry soil. Mineralization was affected by the fertilizer additions ($p=0.001$). It remained lowest in the mineral fertilizer treatment. The deep stable manure resulted in a mineralization which was considerably higher compared to mineral fertilizer (+70%) followed by nature compost (+58%) and the household compost&slurry combination (+39%). The deep stable manure resulted in the highest mineralization followed by nature compost and the household compost&slurry combination.

The overall mineralizable nitrogen did not change between 2006 and 2016 but differences were much more pronounced after 17 years. In 2006 no significant differences were observed between the fertilizer treatments (Koopmans et al., 2006).

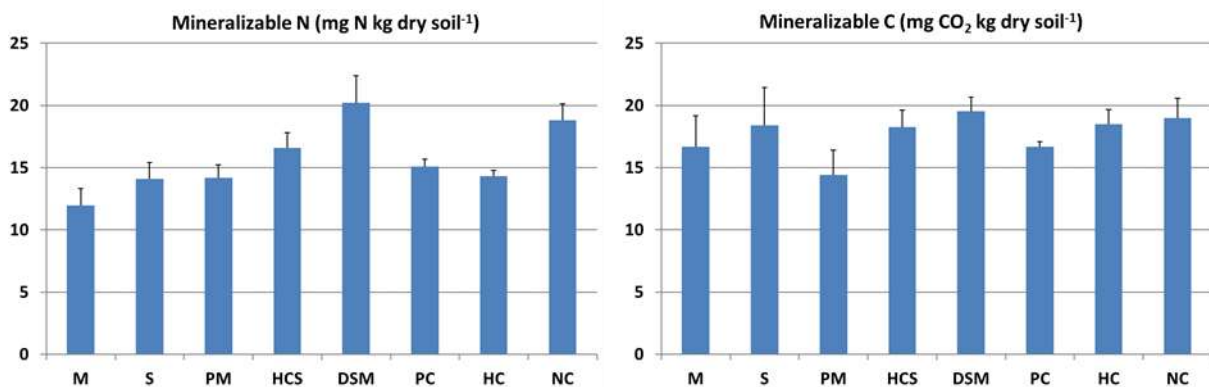


Figure 5: Effect of fertilization treatment on N mineralization (left) and soil mineralizable C (right) for mineral fertilizer (M), slurry (S), poultry manure (PM), household compost&slurry (HCS), deep stable manure (DSM), plant compost (PC), household compost (HC) and nature compost (NC) treatments.

The mineralizable C was on average low at a level of $17.7 \pm 3.5 \text{ mg CO}_2 \text{ 100 g}^{-1} \text{ wk}^{-1}$. Mineralizable C was lowest for poultry manure and highest for deep stable manure. The soil mineralizable C values followed a similar pattern as the level of soil organic matter and were significantly different ($p=0.04$).

The number of earthworms was on average 20 per m^{-2} . Highest numbers were found in the poultry manure treatment followed by the deep stable manure and nature compost treatment, respectively. Lowest numbers were found in the mineral fertilizer treatment and the plant compost treatment. All earthworms found belonged to the endogeic species *Aporrectodea caliginosa* (73%) and *A. chlorotica* (27%).

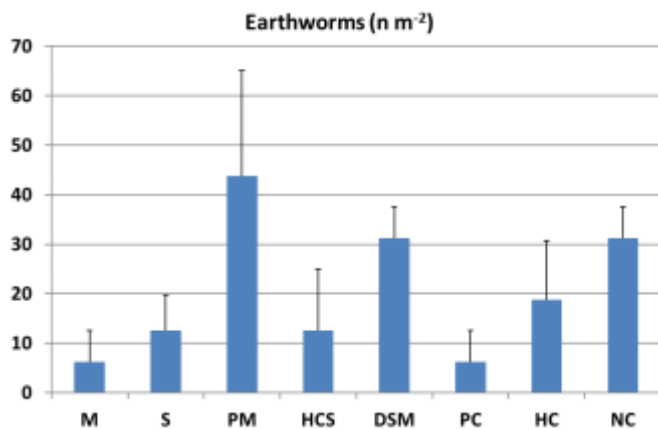


Figure 6: Effect of fertilization treatment on earthworm numbers for mineral fertilizer (M), slurry (S), poultry manure (PM), household compost&slurry (HCS), deep stable manure (DSM), plant compost (PC), household compost (HC) and nature compost (NC) treatments.

Bacterial biomass was between 15 and 27 $\mu\text{g C g}^{-1}$ dry soil which is low compared to averages of 88 and 66 $\mu\text{g C g}^{-1}$ dry soil reported for sand and clay under arable management (Rutgers et al., 2007). Deep stable manure and household compost were at highest levels but the overall bacterial biomass did not differ between fertilizer treatments.

Fungal biomass differed between treatments ($p=0.003$) and was significantly higher in the mineral fertilization treatment compared to all other treatments. In general, fungal biomass however, was low. Active fungi did not differ between the fertilization treatments. The low mineralization in the mineral fertilizer treatment combined with a high fungal biomass could be an indication of a low nutrient availability at the end of the growing season and possibly higher amounts of mycorrhiza fungal hyphae.

With the higher fungal biomass, the fungi/bacteria ratio was found to be highest in the mineral fertilizer treatment. All other treatments remained at levels not significantly different from each other.

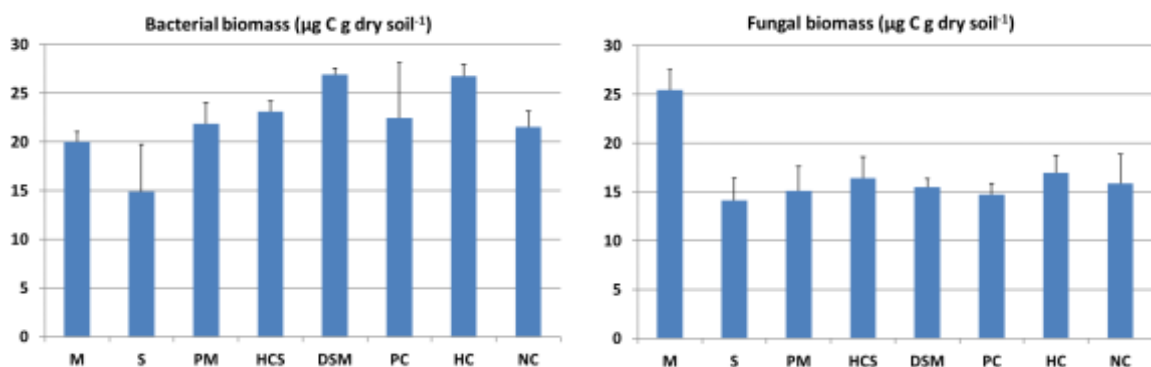


Figure 7: Effect of fertilization treatment on bacterial (left) and fungal (right) biomass for mineral fertilizer (M), slurry (S), poultry manure (PM), household compost&slurry (HCS), deep stable manure (DSM), plant compost (PC), household compost (HC) and nature compost (NC) treatments.

The abundance of nematodes in the fertilization treatments was on average 990 nematodes per 100 g soil^{-1} . This is rather low compared to averages of 1270 and 3605 for clay respectively sand found by Rutgers et al. (2007). Highest absolute numbers were found in the deep stable manure and nature compost followed by mineral fertilizer and plant compost. Herbivorous and bacterivorous nematodes dominated with highest numbers of herbivorous nematodes found in the nature

compost followed by mineral fertilizers and plant compost ($p= 0.088$). No significant differences were found for the smaller numbers of fungivorous nematodes with highest numbers in the household and nature compost ($p=0.099$).

The life-strategy group distribution showed high numbers of cp-1 (enrichment opportunists) in deep stable manure and household compost&slurry. Highest numbers of cp-3 were found for nature compost followed by deep stable manure and household compost with and without slurry. The Maturity Index did not differ between fertilization treatments.

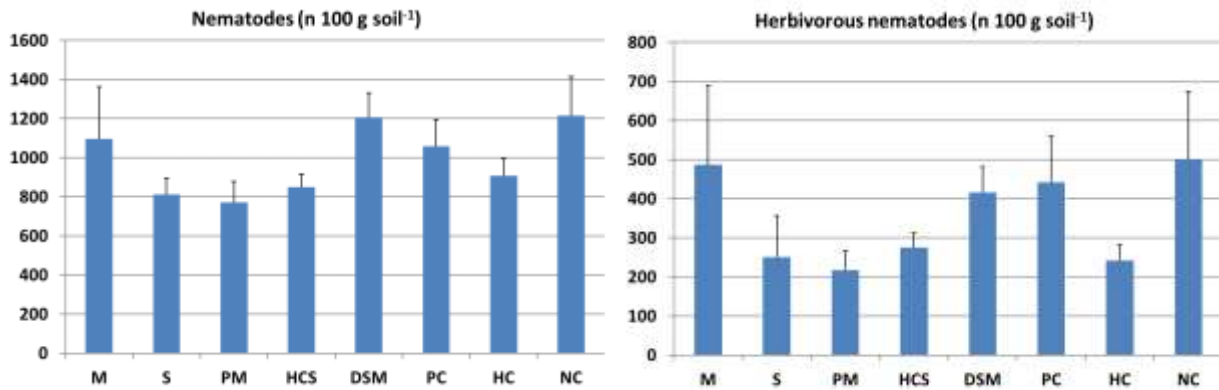


Figure 8: Effect of fertilization treatment on nematode abundance (left) and herbivorous nematodes (right) for mineral fertilizer (M), slurry (S), poultry manure (PM), household compost&slurry (HCS), deep stable manure (DSM), plant compost (PC), household compost (HC) and nature compost (NC) treatments.

4 Discussion and conclusion

In this study, we evaluated the effect of eight organic fertilization treatments on soil quality parameters after 17 years. The experiment was set up in 1999 with certain regulations on inputs. For scientific reasons the fertilization treatments were kept constant during the time of the experiment although regulation changed in the meantime, more in particular, on the P_2O_5 additions allowed in arable rotations.

The treatments have multiple impacts on nutrient and organic matter levels applied in the different strategies. Therefore, the comparison must be seen as 'system comparisons' rather than a one- or multi factor fertilizer study.

Effects on soil organic matter

Different fertilizer treatments resulted in significant differences in soil organic matter levels (Figure 9). To what extent are the resulting organic matter levels due to the fresh organic matter inputs?

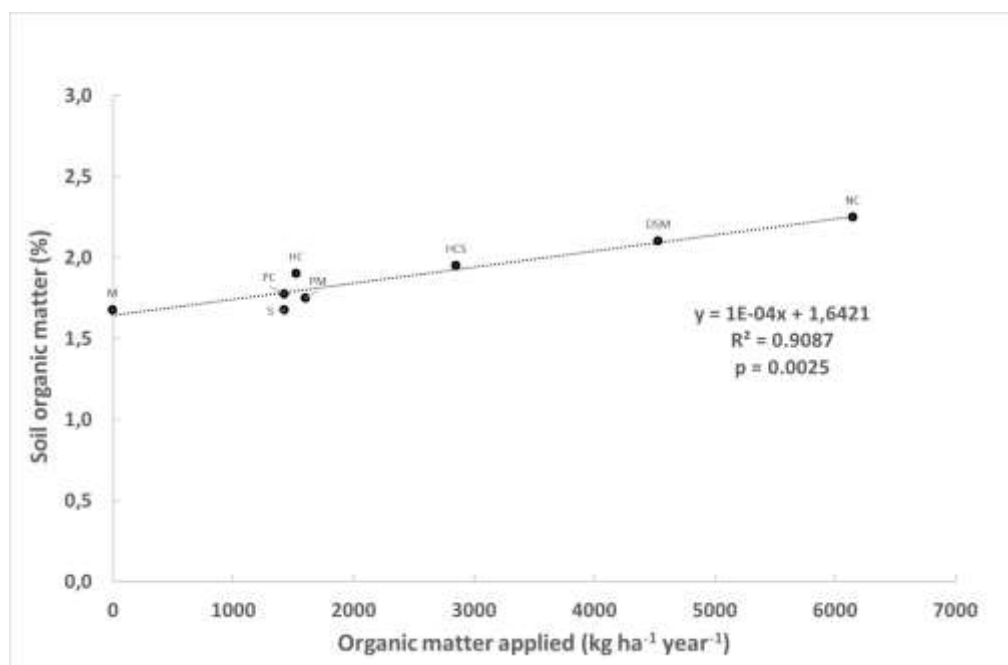


Figure 9: Correlation between amount of fresh organic matter applied in the treatments and organic matter in the soil after 17 years.

Figure 9 shows a high and significant correlation between organic matter levels found in the soil and the amount of fresh organic matter applied to different treatments. The almost straight line suggests that the amount applied is more important than its quality.

Of all treatments focusing on optimal crop nutrition (mineral fertilizer, slurry, poultry manure, household compost&slurry and deep stable manure), the household compost&slurry and deep stable manure treatments include high fresh organic matter additions. This correlates with a build-up of organic matter in the soil.

Of the soil feeding strategies, nature compost, (due to its composition with low nutrient availability not restricted in the amount to be applied to the soil), performed well in terms of fresh organic matter addition and resulting organic matter levels in the soil. The high supply not only resulted in

high organic matter levels but seemed to effect also other soil chemical, physical and biological parameters.

Of special interest are treatments that are well above or below the correlation line. Although more organic matter is applied with the slurry as compared to the mineral fertilizer, after 17 years resulting soil organic matter levels are very similar. Apparently, the organic matter in the slurry did not result in an additional build-up of organic matter. A similar effect was seen when slurry is applied to the household compost. Although this increased the freshly added organic matter by 87% compared to the household compost alone, the increase in soil organic matter in the soil was limited to 0.05%.

Correlation between the yield of the salsify in 2016 and fresh organic matter applied showed no correlation (Appendix 4). Neither correlated the yield with the organic matter level in the soil (Appendix 5).

Effects on potential nitrogen mineralization and yield

Different strategies are associated with different levels of N input. To what extent are resulting yield differences correlated with the actual N application rate and resulting potential soil N mineralization?

Appendix 6 shows there is a very weak correlation between N application and yield. Especially poultry manure resulted in a yield which is above the average expected from the average N application. Results show that yield increased only very slightly due to a higher total N application.

Treatments were designed at similar nitrogen supply expected from mineral N content and mineralization of slurry, household compost&slurry, deep stable manure and mineral fertilizer. Treatments received fertilizers in two out of three years at an intended mineralization potential of 100 kg N per ha per year for the first year resulting in an average of 67 kg N per ha per year (over 3 years) from freshly applied fertilizers. The resulting higher yields we find in slurry, household compost&slurry and deep stable manure as compared to mineral fertilizer alone, might be the result of a build-up of the soil organic matter and especially a build-up of soil N mineralization over time.

Figure 10 indicates that this correlation between N application in organic additions and build-up of soil N mineralization over time indeed exists. In fact, the difference between mineralizable N found in the mineral treatment and the mineralizable N found in slurry or household compost&slurry or deep stable manure can be attributed to residual and cumulative N mineralization from added organic matter after the first year of application. This effect might be referred to as a build-up of soil fertility.

Plant compost and household compost realize a similar build-up of organic matter from similar dry matter additions of 6000 kg per year but slightly differed in N application rate. After 17 years of additions, both treatments did not reach a stage in which soil N mineralization was increased considerably. Consequently, it may be concluded that these soil-building strategies are not increasing mineralization and therefore, the potential for leaching losses due to mineralization, is not increased.

From a farmers perspective however, these soil-building fertilizers neither outperform very much in terms of yield (Appendix 7). As indicated, there is no correlation between the potential soil N mineralization and yield, suggesting there are other factors building-up soil fertility, which might

explain yield differences between treatments. Of interest is the effect on yield of the household compost&slurry treatments with shows highest yields although mineralization is close to the average of all treatments.

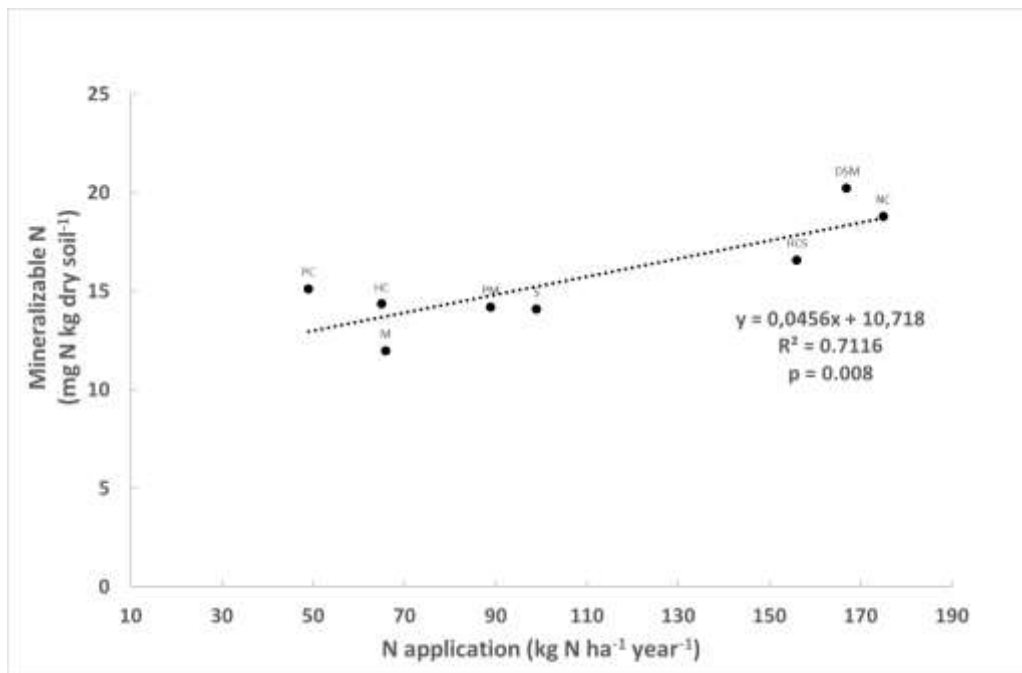


Figure 10: Correlation between mineralizable N and N application.

Although total phosphate application differed considerably between treatments no measurable differences in total phosphate of the soil occurred. A correlation was found between total P_2O_5 application and yield (Appendix 8). Mineral fertilizer and nature compost resulted in below average yields. Poultry manure and nature compost, receiving both 80 kg P_2O_5 per ha per year resulted in yield differences in favour of poultry manure.

Phosphate application also correlated with plant available P, with highest levels found in the deep stable manure. Plant available P could explain 82% of yield differences found in the treatments indicating the system to be very sensitive to plant available P. The soil is likely to be phosphate fixing.

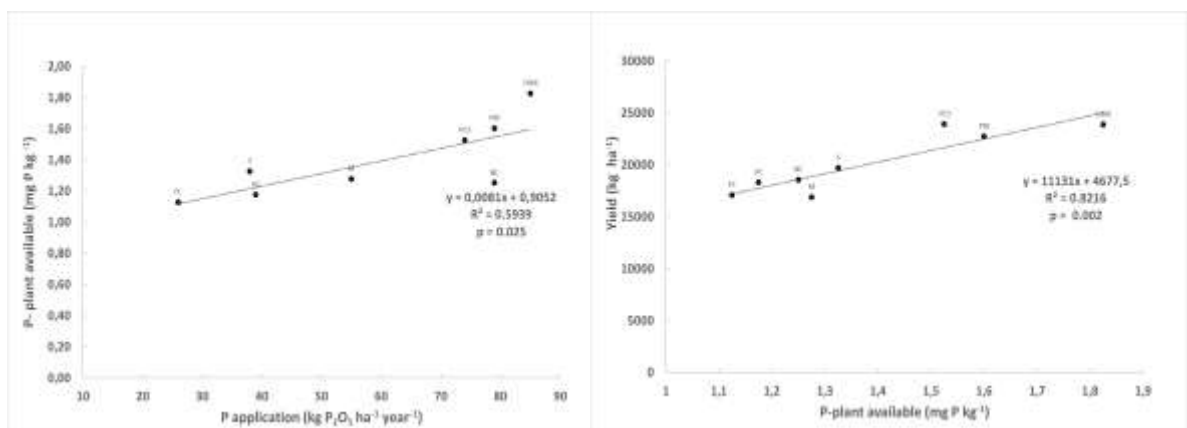


Figure 11: Correlation between plant available P (P-plant) and P application (left) and yield and plant available P (right).

Effects on soil life

HWC shows a relatively strong correlation ($p=0.01$; $R^2=0.69$) with organic matter levels in the soil (Appendix 9). Nature compost resulted in the highest soil organic matter levels of all treatments but HWC was lower as compared to household compost&slurry and deep stable manure. In addition, plant compost and household compost were below the average expected from the correlation.

Higher HWC levels are known to be closely linked to microbial activity. Higher HWC values indicate more soil life and increased soil fertility, although the compounds are less clear than those of mineralizable nitrogen. HWC consists largely of mucus secreted by bacteria and fungi and might cause aggregation. HWC correlated well with total organic matter. From other trials, it is known that it shows faster and greater differences from, for example, organic fertilization or reduced soil cultivation (Bloem et al., 2017). HWC levels neither correlated with number of earthworms ($p=0.34$; $R^2=0.15$) nor soil mineralizable C ($p=0.10$; $R^2=0.38$).

The effect of fertilizers and composts did not result in measureable differences in soil physical parameters. This is not surprising since soil tillage may easily effect these parameters in the topsoil.

Poultry manure seemed to have highest numbers of earthworms per m^{-2} and about 7 times the numbers found in the mineral fertilizer treatment. These earthworms might be important for the root and air penetration of deeper soil layers and drainage.

Figure 12 shows the correlation between the number of earthworms and the amount of macro pores found at 10 and 20 cm soil depth. The correlations seem to confirm that earthworms open up these soil layers through an increased number of earthworm burrows or macro pores.

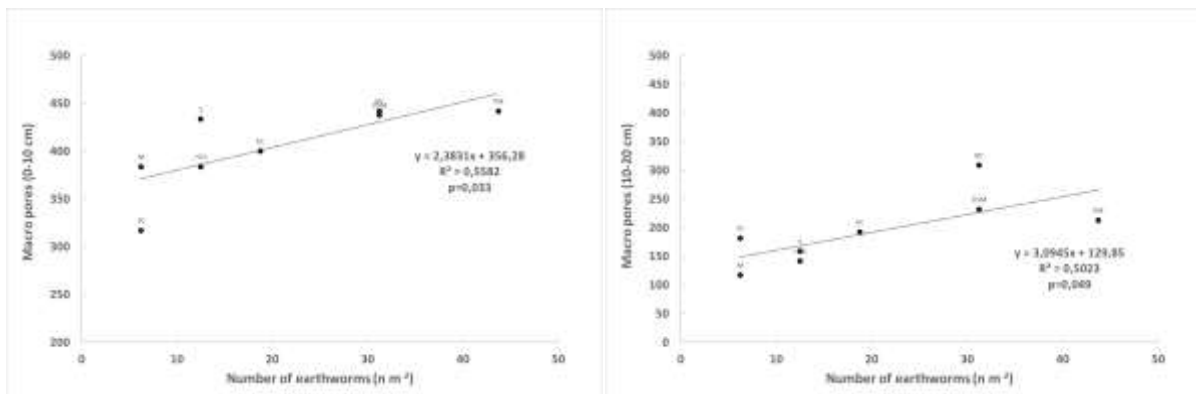


Figure 12: Correlation between number of earthworms and earthworm macro pores found in the soil at 0-10 cm depth (left) and 10-20 cm depth (right).

After 17 years, only limited effects from fertilizer additions were found on bacterial and fungal parameters. Mineralizable C, tended to be highest in the deep stable manure but variability was high. Bacterial biomass found in this soil was low but comparable to other findings for arable land in the Netherlands and Belgium sampled in autumn (Bloem et al., 2006; Koopmans et al., 2006; Van Eekeren et al., 2008). Measurements for the Dutch Soil Quality Monitoring Network (Rutgers et al., 2007) were conducted in spring, shortly after a fertilizer addition and reported higher microbial biomass levels. In our trial, fertilizers were applied in spring of 2015, resulting in a period of over 18 months between the last fertilizer addition and measurement of soil quality. It is likely that bacterial biomass, but also nematode numbers, will be higher shortly after the addition of decomposing fertilizers than at the end of the growing season. We expect our results therefore to be the more long lasting effects of fertilizer additions on soil quality rather than the effect of fresh fertilizer ad-

ditions that may interfere with soil quality. It may explain why some soil life effects are less pronounced.

It is striking that fungal biomass was found to be higher in the mineral fertilizer treatment. In 2006, we found a similar effect, although at that time differences were not significant. In general, fungal biomass is found to increase in lignin rich woody, compost like fertilizer additions. Our results may suggest that low nutrient levels, found in the mineral fertilizer treatment, possibly resulted in higher mycorrhiza levels that may have increased the overall fungal biomass measured.

Nematode abundance was highest in the deep stable manure and nature compost treatments, followed by the mineral fertilizer and plant compost treatment. Levels were about 50% higher if compared to the poultry manure treatment, which showed lowest numbers of nematodes. Most fungivorous nematodes were found in the nature compost and household compost treatments. In the trophic group distribution herbivorous nematodes dominated in the mineral treatment, the plant based compost and the nature compost (> 40% of total). In all other treatments bacterivorous nematodes dominated at levels of 35-42%, which is common for arable, land (Van Eekeren, 2008). The relatively low total numbers of nematodes relative to the Dutch Soil Quality Monitoring Network (Rutgers et al., 2007) might again be related to the time of sampling in November. Especially quick release fertilizers like mineral fertilizer and slurry may cause peaks shortly after addition in spring, but the numbers may have decreased again in the course of the year and especially in autumn if roots quickly decompose. Nature compost, with high additions of wood-like materials may have caused higher numbers of fungivorous nematodes in this treatment.

In the life history groups (cp: colonizer-persister groups), higher levels of cp-1 were found in the household compost&slurry and deep stable manure treatments. The high number of cp-1 indicates a food rich habitat (enrichment opportunists). The typical colonizer family Rhabditidae dominates this life history group. The genera of Tylenchidae and Eucephalobus dominate the cp-2 group of nematodes, which indicate a disturbed habitat.

It may be concluded that effects on soil life are limited. Earthworms show significant differences especially with poultry manure. Nematodes show differences for CP1 and CP3 but no significant effects on bacterial biomass were found. Fungal biomass was found to be highest in mineral fertilizer (applied 18 months before) but effects from organic inputs could not be detected.

After 17 years, significant effects were found for organic matter and labile organic matter levels in the soil. Highest levels of organic matter (+41%) were found in the nature compost treatment followed by deep stable manure (+31%) and household compost&slurry (+22%). The deep stable manure resulted in an N mineralization, which was considerably higher if compared to mineral fertilizer (+70%) followed by nature compost (+58%) and the household compost&slurry combination (+39%). The light soil and intensive soil preparation might give a rapid breakdown and slow build-up of organic matter. Minimizing soil cultivation could still speed up the process of building up organic matter levels in this soil with organic inputs.

Overall, it might be concluded that best results are obtained with fertilizers used for both crop nutrition and building soil quality: poultry manure, household compost&slurry and deep stable manure. Although there are positive indications on total C and labile organic matter, the build-up of the soil and soil organisms is slow which might be the result from this light soil and the use of intensive soil cultivation.

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Appendix 1

Soil Chemical and physical characteristics in 2016 of mineral (M), slurry (S), poultry manure (PM), household compost&slurry (HCS), deep stable manure (DSM), plant compost (PC), household compost (HC) and nature compost (NC).

	Unit	Treatments								P-value
		M	S	PM	HCS	DSM	PC	HC	NC	
pH-KCL	-log[H ⁺]	7,80 ab	7,80 ab	7,80 ab	7,78 ab	7,75 a	7,83 b	7,80 b	7,75 a	0,082
SOM	%	1,68 a	1,68 a	1,75 a	1,95 bc	2,10 cd	1,78 ab	1,90 ab	2,25 d	< 0,001
HWC	ug C g ⁻¹	277 a	291 a	293 a	357 b	376 b	278 a	284 a	355 b	0,003
C-total	g C kg dry soil ⁻¹	14,5 abc	13,3 a	13,5 a	14,0 ab	15,5 bc	13,5 a	14,5 ab	16,0 c	0,005
N-total	g N kg dry soil ⁻¹	1,11 a	1,09 a	1,13 ab	1,19 abc	1,28 bc	1,09 a	1,18 abc	1,29 c	0,018
P-total	mg P ₂ O ₅ 100 g ⁻¹	116	108	118	111	109	108	107	108	NS
P-plant	mg P kg ⁻¹	1,28 a	1,33 ab	1,60 cd	1,53 bc	1,83 d	1,13 a	1,18 a	1,25 a	< 0,001
K	mg K kg ⁻¹	65 bcd	63 aabc	56 ab	75 d	89 e	55 a	53 a	67 cd	< 0,001
Ca	mmol Ca L ⁻¹	0,53	0,65	0,73	0,58	0,78	0,65	0,50	0,45	NS
Mg	mg Mg kg ⁻¹	29,75 a	32,5 a	36,75 b	38,25 b	38,5 b	32 a	32,5 a	39,25 b	< 0,001
Bulk density	g cm ⁻³	1,26	1,24	1,28	1,23	1,22	1,30	1,25	1,20	NS
Soil structure 0-10 cm										
Crumb	%	74	65	70	74	64	66	66	60	NS
Sub-angular	%	24	31	23	26	33	31	29	34	NS
Angular	%	3	4	8	0	4	3	5	6	NS
Soil structure 10-20 cm										
Crumb	%	50	46	56	34	51	38	33	54	NS
Sub-angular	%	44	39	38	56	43	43	46	39	NS
Angular	%	6	15	6	10	6	20	21	8	NS
Macro pores 0-10	#	383	433	442	383	438	317	400	442	NS
Macro pores 0-20	#	117	158	213	142	231	181	192	308	NS

Appendix 2

Soil biological characteristics in 2016 of mineral (M), slurry (S), poultry manure (PM), household compost&slurry (HCS), deep stable manure (DSM), plant compost (PC), household compost (HC) and nature compost (NC).

	Unit	Treatments								P-value
		M	S	PM	HCS	DSM	PC	HC	NC	
Mineralizable N	mg N kg dry soil ⁻¹	11,9 a	14,1 ab	14,2 ab	16,6 bc	20,2 d	15,1 ab	14,3 ab	18,8 cd	0,001
Mineralizable C	mg CO ₂ kg dry soil ⁻¹	16,7 ab	18,4 ab	14,4 ab	18,3 ab	19,5 b	16,7 ab	18,5 ab	19,0 b	0,039
Number of earthworms	n m ⁻²	6 a	13 ab	44 b	13 ab	31 ab	6 ab	19 ab	31 ab	0,021
Bacterial biomass	µg C g dry soil ⁻¹	20,0	14,9	21,9	23,1	26,9	22,5	26,8	21,5	NS
Fungal biomass	µg C g dry soil ⁻¹	25,4 b	14,2 a	15,1 a	16,4 a	15,5 a	14,7 a	17,0 a	15,9 a	0,033
Active fungi	µg C g dry soil ⁻¹	1,09	2,40	1,87	1,10	0,75	0,86	1,29	2,37	NS
Fungi/Bacteria	C C ⁻¹	1,27 b	1,49 a	0,73 a	0,72 a	0,58 a	0,83 a	0,64 a	0,73 a	0,033

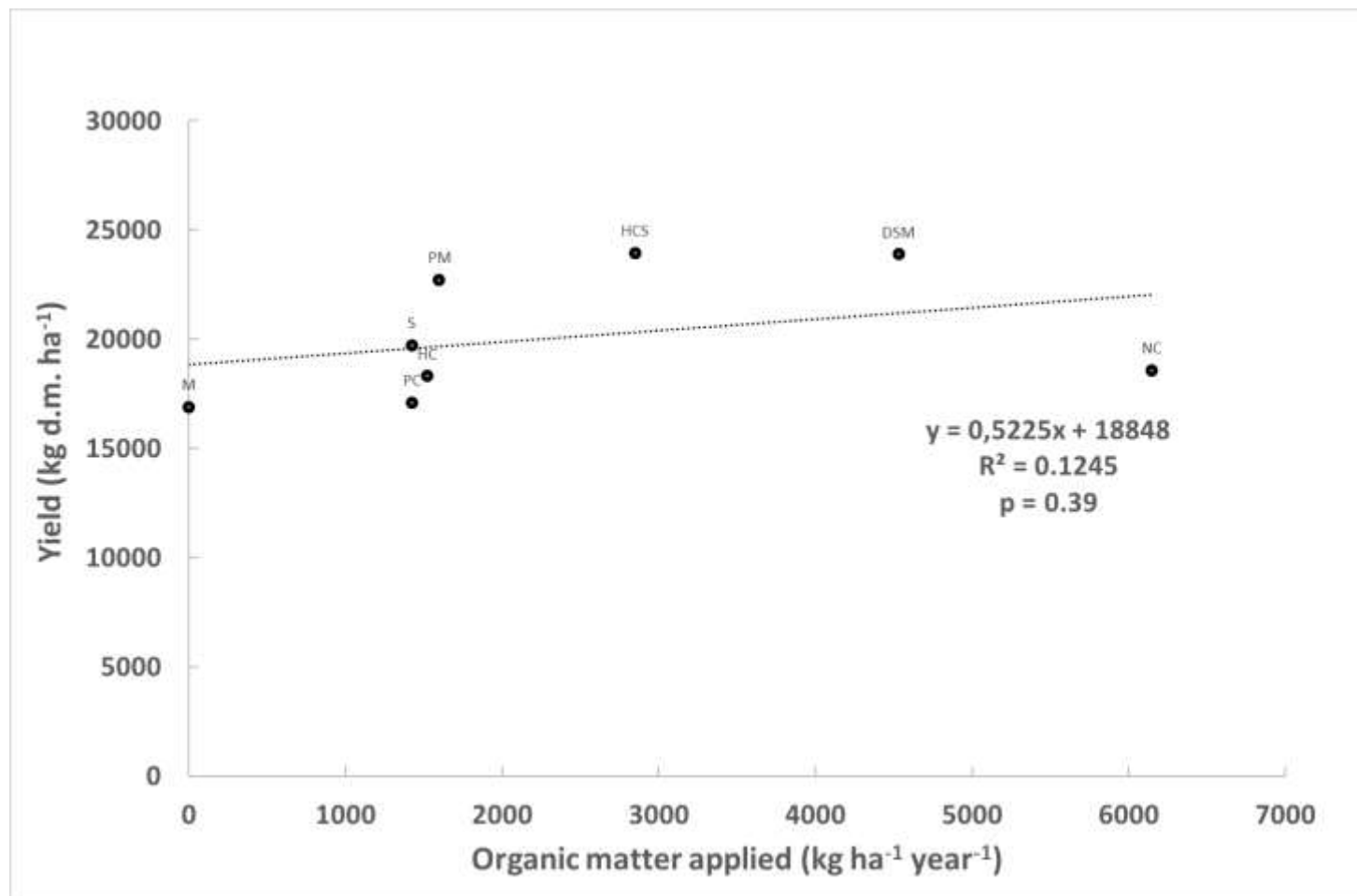
Appendix 3

Nematode abundance, trophic groups, life history groups (cp colonizer-persister groups) and cp: community structure indices in 2016 of mineral (M), slurry (S), poultry manure (PM), household compost&slurry (HCS), deep stable manure (DSM), plant compost (PC), household compost (HC) and nature compost (NC).

	Unit	Treatments								P-value
		M	S	PM	HCS	DSM	PC	HC	NC	
Number of nematodes	n 100 g soil ⁻¹	1097 abc	812 ab	772 a	849 abc	1208 bc	1058 abc	908 abc	1214 c	0,088
Bacterivorous	n 100 g soil ⁻¹	335	286	276	308	454	308	379	403	NS
Fungivorous	n 100 g soil ⁻¹	72 abc	70 ab	69 a	82 ab	78 ab	74 abc	121 ab	126 b	0,099
Herbivorous	n 100 g soil ⁻¹	488	250	218	275	416	442	243	502	NS
Carnivorous	n 100 g soil ⁻¹	199	199	206	182	257	232	159	182	NS
General	n 100 g soil ⁻¹	4	7	3	2	2	3	6	1	NS
cp-1	%	75,6 abc	81,9 abc	80,5 abc	119,3 bc	124,4 c	52,6 a	72,8 ab	68,8 ab	0,017
cp-2	%	646,7	347,1	301,1	290,8	546,4	528,3	394,3	605,8	NS
cp-3	%	137,5 a	138,3 a	145,9 ab	222,6 c	230,4 c	215,7 bc	225,4 bc	301,2 d	0,001
cp-4	%	137,8	157,4	179,8	152,6	220,1	186,5	148,5	179,1	NS
cp-5	%	94,8	80,5	61,4	62,1	83,9	72,0	61,4	57,7	NS
Maturity index (cp1-5)	µg C g dry soil ⁻¹	3,00	3,07	2,97	2,85	2,89	3,22	2,88	2,98	NS
Maturity index (cp2-5)	µg C g dry soil ⁻¹	3,24	3,44	3,31	3,35	3,24	3,39	3,11	3,19	NS

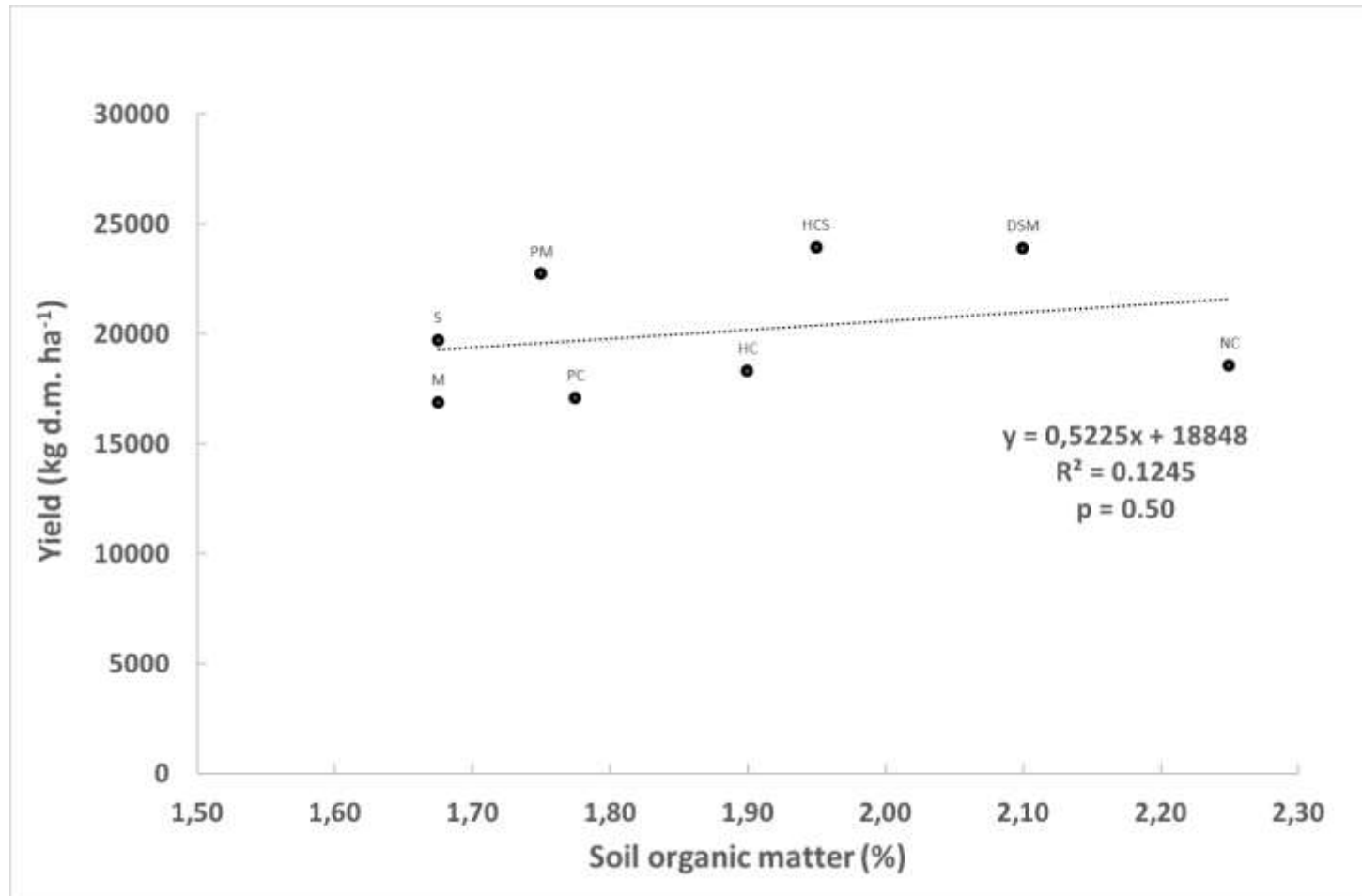
Appendix 4

Correlation between yield and organic matter applied for mineral fertilizer (M), slurry (S), poultry manure (PM), household compost&slurry (HCS), deep stable manure (DSM), plant compost (PC), household compost (HC) and nature compost (NC) treatments after 17 years.



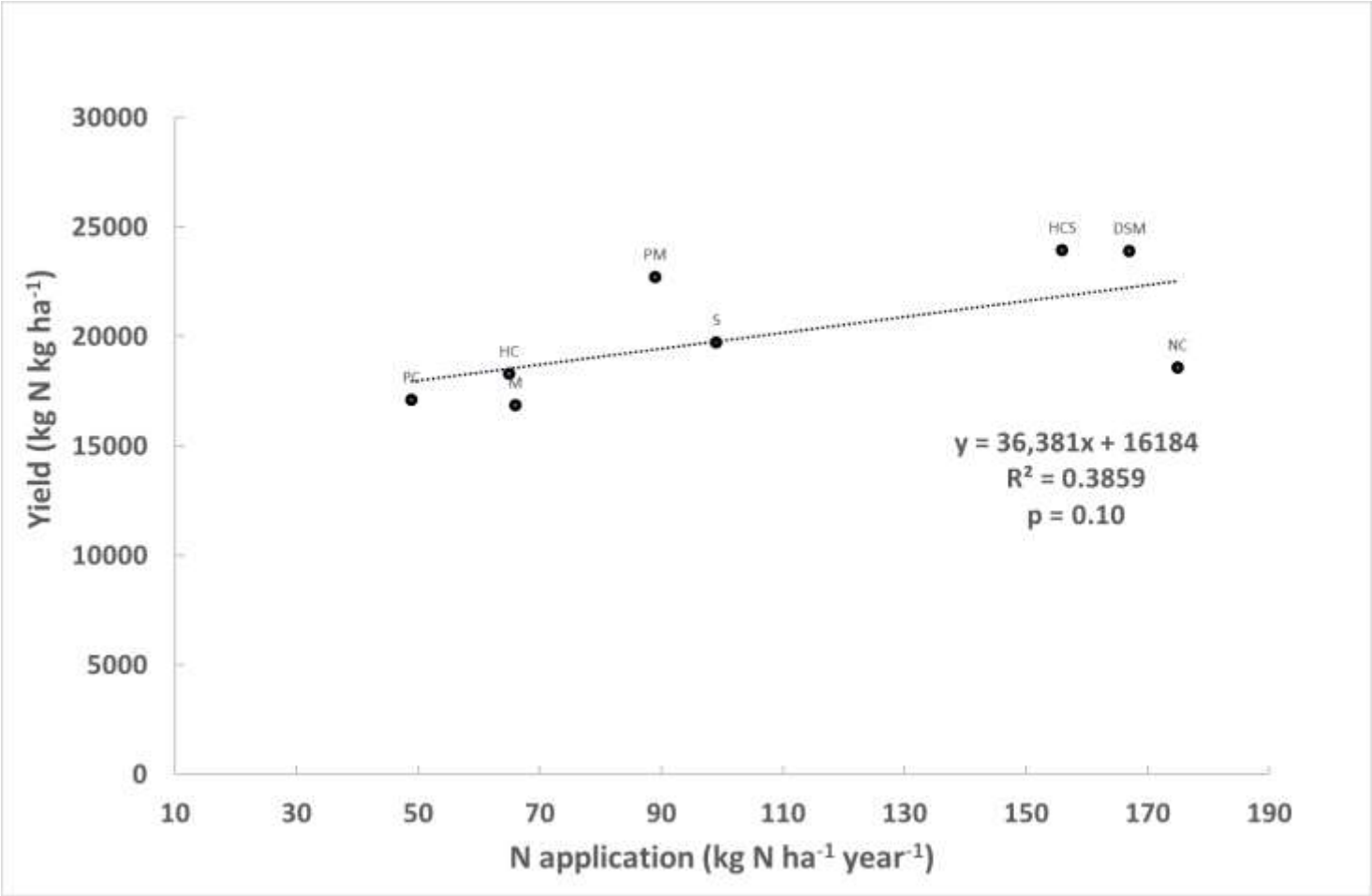
Appendix 5

Correlation between yield and soil organic matter for mineral fertilizer (M), slurry (S), poultry manure (PM), household compost&slurry (HCS), deep stable manure (DSM), plant compost (PC), household compost (HC) and nature compost (NC) treatments after 17 years.



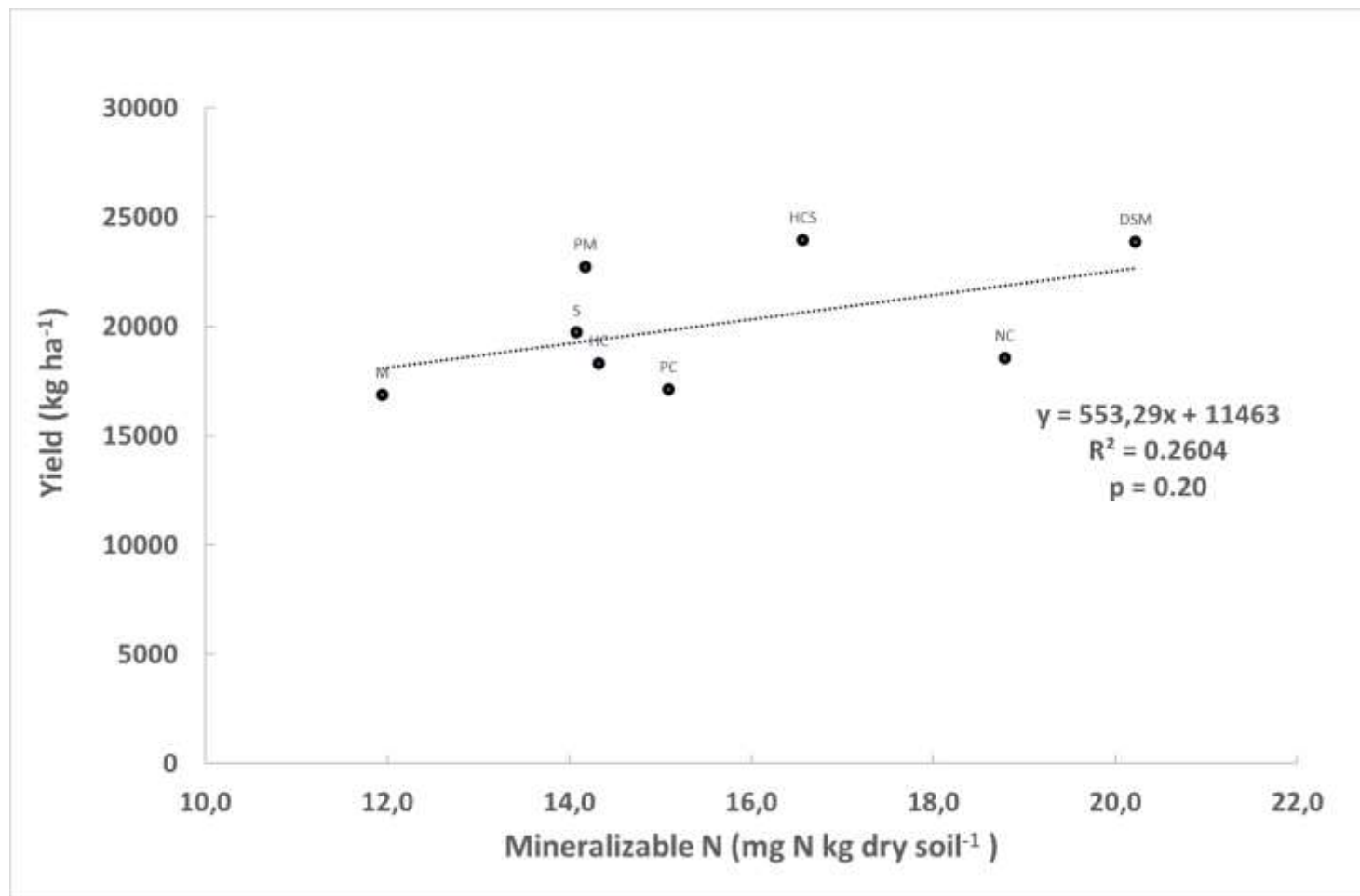
Appendix 6

Correlation between yield and N application for mineral fertilizer (M), slurry (S), poultry manure (PM), household compost&slurry (HCS), deep stable manure (DSM), plant compost (PC), household compost (HC) and nature compost (NC) treatments after 17 years.



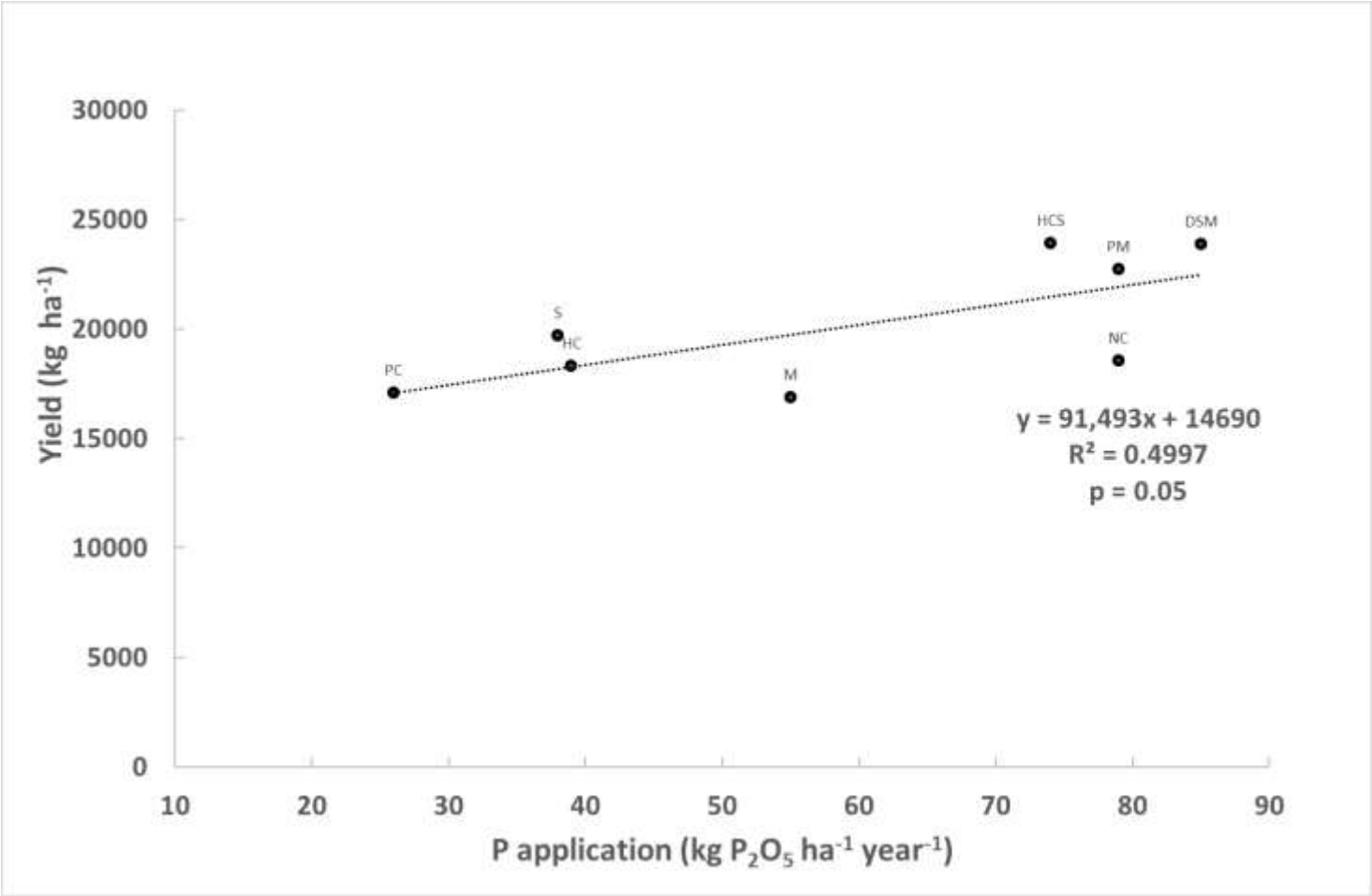
Appendix 7

Correlation between yield and mineralizable N for mineral fertilizer (M), slurry (S), poultry manure (PM), household compost&slurry (HCS), deep stable manure (DSM), plant compost (PC), household compost (HC) and nature compost (NC) treatments after 17 years.



Appendix 8

Correlation between yield and P application for mineral fertilizer (M), slurry (S), poultry manure (PM), household compost&slurry (HCS), deep stable manure (DSM), plant compost (PC), household compost (HC) and nature compost (NC) treatments after 17 years.



Appendix 9

Correlation between yield and soil organic matter for mineral fertilizer (M), slurry (S), poultry manure (PM), household compost&slurry (HCS), deep stable manure (DSM), plant compost (PC), household compost (HC) and nature compost (NC) treatments after 17 years.

